

Distribution and genetic diversity of two species of *Pelagus* minnows (Leuciscidae) in southern Greece

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Abstract – *Pelagus* Kottelat & Freyhof 2007 (Leuciscidae) is a freshwater fish genus endemic to the Southern Balkans. The distribution of most of its species is insufficiently known. *Pelagus* was molecularly studied only marginally, and the genetic diversity of individual species or populations, crucial for their conservation, is completely unknown. We studied distribution and genetic diversity of *Pelagus stymphalicus* (Valenciennes 1844) and *Pelagus marathonicus* (Vinciguerra 1921), two widespread species from southern Greece. Our data, based on *cytochrome b* sequences, confirmed that a number of populations whose taxonomic status had been uncertain, belong to one of these species. The distribution range of *P. stymphalicus* includes the Ionian mainland, from the Acheloos to Mornos rivers, and most of the Peloponnese, excluding the Evrotas and the headwaters of Alfios. The native range of *Pelagus marathonicus* is the western Aegean mainland, from rivers near Athens to the Xerias river in the Pagasitikos Gulf. AMOVA showed that the genetic variance is slightly higher among than between populations. High F_{ST} values indicated a pronounced genetic differentiation of the populations in both species. There was a high proportion of private haplotypes and a very small number of shared haplotypes between populations in both species, which indicates uniqueness of each population and their susceptibility to human-induced changes.

Keywords: cytochrome b / *Pelagus stymphalicus* / *Pelagus marathonicus* / Balkans / endemic species

Résumé – Répartition et diversité génétique de deux espèces de vairons *Pelagus* (Leuciscidae) dans le sud de la Grèce. *Pelagus* (Kottelat & Freyhof 2007) (Leuciscidae) est un genre de poisson d'eau douce endémique des Balkans du Sud. La distribution de la plupart de ses espèces est insuffisamment connue. Le genre *Pelagus* n'a fait l'objet que d'une étude moléculaire marginale, et la diversité génétique des espèces ou des populations, cruciale pour leur conservation, est totalement inconnue. Nous avons étudié la répartition et la diversité génétique de *Pelagus stymphalicus* (Valenciennes 1844) et de *Pelagus marathonicus* (Vinciguerra 1921), deux espèces très répandues dans le sud de la Grèce. Nos données, basées sur les séquences du cytochrome b, ont confirmé qu'un certain nombre de populations dont le statut taxonomique était incertain, appartiennent à l'une de ces espèces. L'aire de répartition de *P. stymphalicus* comprend le continent ionien, des rivières Acheloos à Mornos, et la plus grande partie du Péloponnèse, à l'exclusion l'Evrotas et l'amont de l'Alfios. L'aire de répartition naturelle du *Pelagus marathonicus* est la partie continentale occidentale de la mer Égée, depuis les rivières près d'Athènes jusqu'à la rivière Xerias dans le golfe de Pagasitikos. L'AMOVA a montré que la variance génétique est légèrement plus élevée au sein des populations qu'entre elles. Les valeurs élevées de F_{ST} indiquent une différenciation génétique prononcée des populations chez les deux espèces. Il y avait une forte proportion d'haplotypes particuliers et un très petit nombre d'haplotypes partagés entre les populations des deux espèces, ce qui indique le caractère unique de chaque population et sa sensibilité aux changements induits par l'homme.

Mots clés : cytochrome b / *Pelagus stymphalicus* / *Pelagus marathonicus* / les Balkans / espèces endémiques

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1 Introduction

The Southern Balkans host distinctive freshwater fish assemblages, which have developed during a long and complex geological history of the area. Several biogeographic freshwater ecoregions have been recognized here, based on the distribution of fish species (Zogaris *et al.*, 2009; Oikonomou *et al.*, 2014). Two of these regions, the Ionian and Western Aegean (Zogaris and Economou, 2017; Vavalidis *et al.*, 2019), lie in southern Greece. These regions substantially differ in their ichthyofaunal characteristics. While the Ionian region is known for its elevated number of species, the Western Aegean is considered species-poor (Economou *et al.*, 2016; Zogaris and Economou, 2017). However, despite a high degree of endemism in both regions, the knowledge on genetic diversity of fish species, and often also on their exact distribution ranges and even biology, is still inadequate (Economidis, 2009; Zogaris and Economou, 2017). Recently conducted detailed genetic studies on a particular fish species have always revealed some unexpected patterns. For example, a high genetic diversity was discovered in the genus *Valencia* Myers 1928 (Vogiatzi *et al.*, 2014; Freyhof *et al.*, 2014), which led to the description of a new species from the Peloponnese and the southwestern mainland (Freyhof *et al.*, 2014). Similarly, a study on *Telestes pleurobipunctatus* (Stephanidis 1939) revealed a high genetic differentiation between populations from different basins, suggesting the existence of a complex of species in the Ionian Sea slope (Buj *et al.*, 2019). On the other hand, *Knipowitschia goernerii* Ahnelt 1991, which was believed to be restricted to a single lagoon system on Kerkira Island, is most probably much more widespread in the Ionian biogeographic region (Vanhove *et al.*, 2016; Vukić *et al.*, 2016, 2017). Finally, molecular analyses of the species of *Squalius* Bonaparte 1837 from the south of the Peloponnese uncovered a past hybridization between evolutionary very distant lineages, which led to a complete mitochondrial DNA introgression (Perea *et al.*, 2016).

Among the most widespread fishes in southern Greece are members of the leuciscid minnow genus *Pelasgus* Kottelat & Freyhof 2007, which is endemic to the southern Balkans. Out of the seven recognized species, three are present in southern Greece (Barbieri *et al.*, 2015). The Evrotas minnow *Pelasgus laconicus* (Kottelat & Barbieri 2004) occurs only in two river basins in the southern Peloponnese (Evrotas and headwaters of Alfios), while the Marathon minnow *Pelasgus marathonicus* (Vinciguerra 1921) and the Stymphalia minnow *Pelasgus stymphalicus* (Valenciennes 1844) are supposed to have large distribution ranges, even though a number of populations were only tentatively identified as belonging to these species (Kottelat and Barbieri, 2004; Kottelat and Freyhof, 2007). While *P. stymphalicus* occurs in a number of drainages in the southern part of the Ionian region (both on the mainland and on the Peloponnese), *P. marathonicus* inhabits rivers and streams of the Western Aegean region (Kottelat and Barbieri, 2004; Economou *et al.*, 2007; Koutsikos *et al.*, 2012; Barbieri *et al.*, 2015). Both species are important biogeographic indicator species for the Western Aegean (*P. marathonicus*) and Ionian (*P. stymphalicus*) ecoregions (Vavalidis *et al.*, 2019) and thus are of a great conservation importance. Nevertheless, no detailed study, genetic nor morphologic, which would include specimens of *P. marathonicus* or *P. stymphalicus* from multiple

localities has been published. The aim of this work is to study the distribution and genetic diversity of populations of *P. marathonicus* and *P. stymphalicus*, based on recent samples comprehensively covering their expected ranges, and on molecular analyses of the mitochondrial marker *cytochrome b*.

2 Material and methods

One-hundred-twenty-five individuals were collected by electrofishing from 29 localities in southern Greece, extensively covering expected distribution ranges of *P. marathonicus* (48 individuals from eight river basins) and *P. stymphalicus* (77 individuals from twelve river basins; Tab. 1, Fig. 1; see Tabs. S1, S2 for details). Finclips were preserved in 96% ethanol for subsequent molecular analysis. Genomic DNA was extracted using the Geneaid™ DNA Isolation Kit (Tissue) (Geneaid Biotech). Individuals were genotyped by sequencing of the mitochondrial marker *cytochrome b* (*cyt b*), which is widely used in population (*e.g.* Pascual *et al.*, 2017; Buj *et al.*, 2019), phylogenetic (*e.g.* Perea *et al.*, 2010; Yang *et al.*, 2015) and taxonomic studies (*e.g.* Buj *et al.*, 2010; Barbieri *et al.*, 2017).

Amplification was performed with a commercial kit (PPP Master Mix; Top-Bio) in a GeneTouch thermocycler (Hangzhou Bioer Technology Co. Ltd.). *Cyt b* was amplified with the primers GluF and ThrR (Machordom and Doadrio, 2001) under the conditions described in Šanda *et al.* (2008).

PCR products were visualized on 1.5% agarose gel and purified by ethanol precipitation (Towner, 1991). Sequencing reactions were performed by MacroGen Service Centre Europe (Amsterdam, Netherlands). The primers used for sequencing were GluF, ThrR (Machordom and Doadrio, 2001), CB4-Glu and PhoxThr (Buj *et al.*, 2019).

Sequences were visually checked and corrected in Chromas v2.6.6 (<http://technelysium.com.au/wp/chromas/>) and BioEdit v7.2.6 (Hall, 1999). The sequences of individual haplotypes were deposited in GenBank under accession numbers MT497285–MT497331. For each species one of the obtained haplotypes was already published (Schönhuth *et al.*, 2018; Benovics *et al.*, 2020; see Tab. S3). Sequences from more localities within the same river basin were pooled together per basin for all analyses. Sequences of individuals from the type localities were taken as a reference and all sequences were compared to the published *cyt b* sequences of *Pelasgus* species (Zardoya and Doadrio, 1999; Freyhof *et al.*, 2006; Perea *et al.*, 2010; Schönhuth *et al.*, 2018; Benovics *et al.*, 2020; see Tab. S3). For this, a Bayesian Inference (BI) tree was constructed (Fig. S1). The appropriate model of nucleotide substitution was selected using jModelTest v2.1.7 (Darriba *et al.*, 2012), based on Akaike Information Criterion (AIC). The selected model was GTR + I. BI was conducted in MrBayes v3.2.2 (Ronquist *et al.*, 2012) with four independent MCMC runs for 3 million generations. Trees were sampled every 600 generations. The convergence of the runs was analyzed and visualized in Tracer v1.7.1 (Rambaut *et al.*, 2018). The first 20% of sampled trees were discarded as burn-in and the remaining trees were used to construct a 50% majority-rule consensus tree. For phylogenetic analysis individual haplotypes were used. The sequences were collapsed to haplotypes in DnaSP v6.12.01

Table 1. Sampling localities, number of analyzed individuals (*N*), number of haplotypes (*h*) and haplotype designation (haplotypes are numbered from 1 for each species).

Species	River basin	No of sampled sites	N	h	Haplotypes
<i>P. marathonicus</i>	Xerias	1	5	2	Hap_1, hap_2
	Kifissos (Beotia)	2	8	5	Hap_3, hap_4 , hap_5, hap_6, hap_7
	Erasinos	1	5	2	Hap_8 , hap_9
	Kato Souli (Marathon)	1	8	3	Hap_10, hap_11, hap_12
	Kifissos (Athens)	1	5	1	Hap_8
	Pinios (Thessaly)	1	5	4	Hap_4 , hap_13 , hap_14, hap_15
	Sperchios	2	4	4	Hap_13 , hap_15 , hap_16, hap_17
	Assopos (Beotia)	2	8	1	Hap_18
<i>P. stymphalicus</i>	Mornos	3	7	4	Hap_1, hap_2, hap_4, hap_5
	Evinos	2	4	2	Hap_3 , hap_6
	Acheloos	1	4	4	Hap_3 , hap_7, hap_8, hap_9
	Piros	1	2	1	Hap_10
	Kotichi	1	4	4	Hap_11, hap_12, hap_13, hap_14
	Pinios	2	6	1	Hap_14
	Alfios	3	9	6	Hap_15, hap_16, hap_17, hap_18, hap_19 , hap_20
	Peristeras	1	2	1	Hap_21
	Pamissos	1	5	2	Hap_19 , hap_22
	Milos	1	3	1	Hap_19
	Assopos (Peloponnese)	1	5	4	Hap_23, hap_24, hap_25 , hap_26
	Stymphalia	2	26	9	Hap_19 , hap_22 , hap_25 , hap_26 , hap_27, hap_28, hap_29, hap_30, hap_31

(Rozas *et al.*, 2017). *Delminichthys ghetaldii* (Steindachner 1882) was used as outgroup. Inkscape v0.92.3 was used for editing the tree.

Intraspecific genetic diversity was estimated by calculating several DNA polymorphism measures in DnaSP v6.12.01 (Rozas *et al.*, 2017): number of polymorphic sites (*S*), nucleotide diversity (π), number of haplotypes (*h*) and haplotype diversity (*Hd*). MEGA v7.0.26 (Kumar *et al.*, 2016) was used to calculate mean within and between population genetic distances (uncorrected p-distances). DnaSP (v6.12.01, Rozas *et al.*, 2017) was further used for tests of neutrality (Tajima's and Fu and Li's tests), for sorting *cyt b* haplotypes and to prepare input files for the Analysis of Molecular Variance (AMOVA). Haplotype networks were constructed in TCS v1.21 (Clement *et al.*, 2000) based on statistical parsimony estimation (Templeton *et al.*, 1992) with a 95% connection limit. Visualization and editing of the haplotype networks were performed in the web-based program tcsBU (TCS Beautifier; Santos *et al.*, 2016) and Inkscape v0.92.3. Analysis of Molecular Variance (AMOVA), including fixation indices (F_{STs}) estimations, was performed in Arlequin v 3.5.2.2. (Excoffier and Lischer, 2010) executing 16,000 permutations. For the estimation of significance of pairwise F_{STs} a standard Bonferroni correction was applied. A number of genetically similar clusters within each species was inferred with Discriminant Analysis of Principal Components (DAPC) implemented in adegenet package v2.1.1 (Jombart, 2008) for R software (v3.6.1). DAPC is a multivariate approach which first transforms the data using a principal component analysis (PCA) and subsequently identifies clusters using discriminant analysis. Best-fit nucleotide substitution model was estimated independently for each species with jModelTest v2.1.7

(Darriba *et al.*, 2012) and selected on the basis of Akaike information criterion (AIC) and Bayesian information criterion (BIC). Selected model (TN93 for both species) was used for inference of population demography applying the coalescent Bayesian Skyline Plot model (Drummond *et al.*, 2005) implemented in BEAST v1.7.5 (Drummond *et al.*, 2012). *Cyt b* evolutionary rate of 0.4% divergence per lineage per million years estimated for leuciscids (Perea *et al.*, 2010) was applied with a strict clock model. Length of MCMC chain was 100 million and sampling every 10,000 generations. Results from three independent runs were combined using LogCombiner v1.7.5 (Drummond and Rambaut, 2007) with a burn-in of 20 million iterations (20%) for each run. Tracer v1.7.1 (Rambaut *et al.*, 2018) was used to assess the convergence of chains and to reconstruct Bayesian Skyline Plots.

3 Results

Forty-eight individuals were identified as *P. marathonicus* and 77 individuals as *P. stymphalicus* (see Tab. 1). The phylogenetic relationships of species of *Pelastgus* are shown in Figure S1. The minimum interspecific genetic distances on *cyt b* between any pair of species of the genus *Pelastgus* was 3.75%, and was found between *P. stymphalicus* and *P. marathonicus* (estimated from own data and published sequences from NCBI database, not shown). The maximum intraspecific differences in both investigated species in this work did not exceed 0.9%. These results show a conspecific status of the investigated populations of each species.

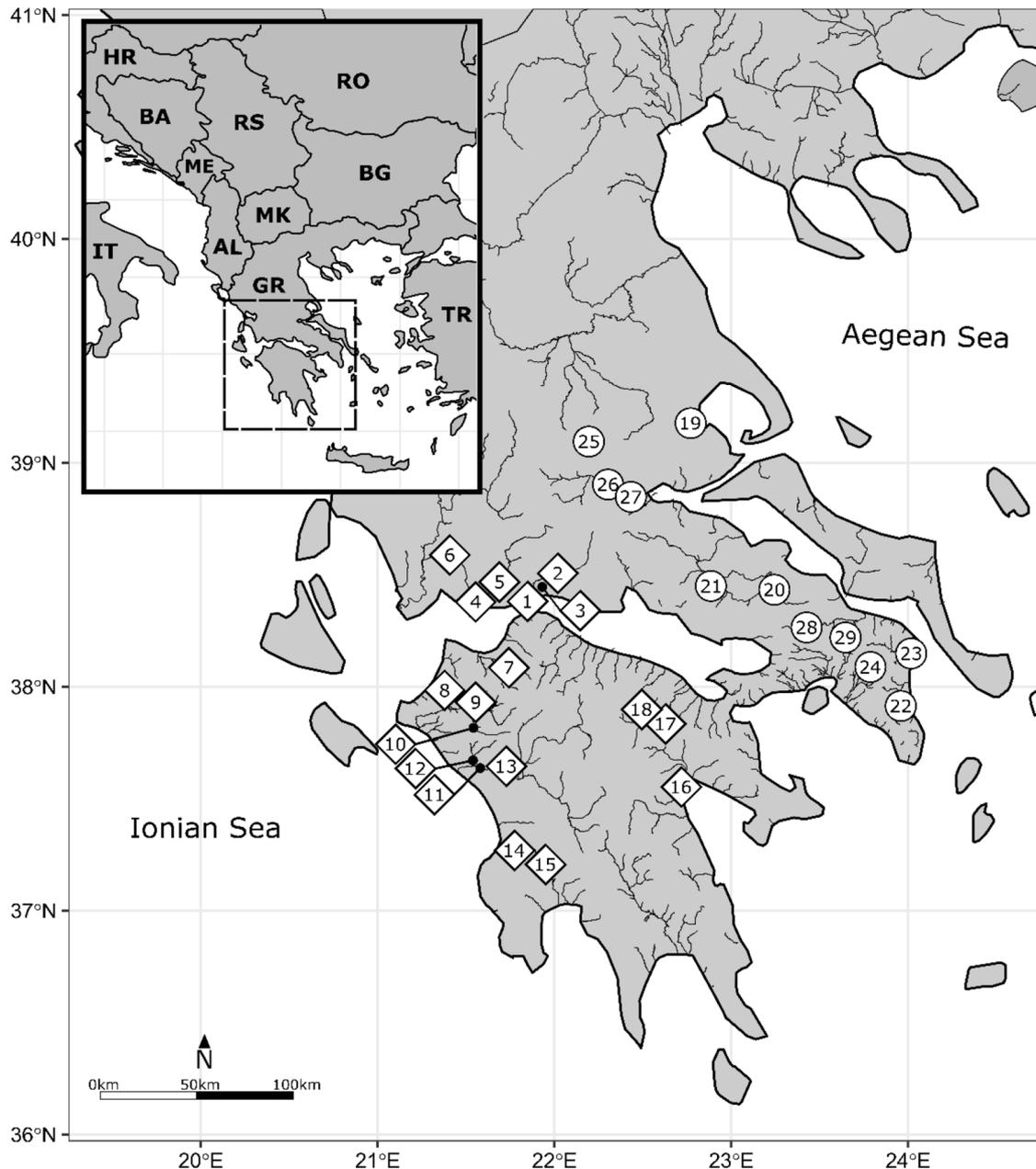


Fig. 1. Sampling sites. *Pelasgus stymphalicus* – \diamond : 1,2,3–Mornos; 4,5–Evinos; 6–Acheloois; 7–Piros; 8–Kotichi Lake; 9,10–Piniot-Peloponnese; 11,12,13–Alfios; 14–Peristeras; 15–Pamissos; 16–Milos; 17–Assopos-Peloponnese; 18–Stymphalia Lake. *Pelasgus marathonicus* – \circ : 19–Xerias; 20,21–Kifissos-Beotia; 22–Erasinos; 23–Kato Souli; 24–Kifissos-Athens; 25–Piniot-Thessaly; 26,27–Sperchios; 28,29–Assopos-Beotia. See Tabs S1, S2 for details.

3.1 *Pelasgus marathonicus*

A total of 48 sequences of *cyt b*, of 1140 bp length, were obtained for *P. marathonicus*. We revealed a high number of *cyt b* haplotypes ($h=18$) and haplotype diversity ($Hd=0.93$; Tab. 2) within this species. The tests of neutrality were all negative, but nonsignificant. Bayesian Skyline Plot reconstruction showed a slightly increasing population size over time, with a recent tendency to decrease, dated to approximately 10 thousand years ago (Fig. 2a).

Table 2. Measures of genetic diversity of *Pelasgus marathonicus* and *P. stymphalicus*. N – number of sequences; π – nucleotide diversity; S – number of polymorphic sites; h – number of haplotypes; Hd – haplotype diversity.

	N	π	S	h	Hd
<i>P. marathonicus</i>	48	0.00297	19	18	0.93
<i>P. stymphalicus</i>	77	0.00370	39	31	0.93

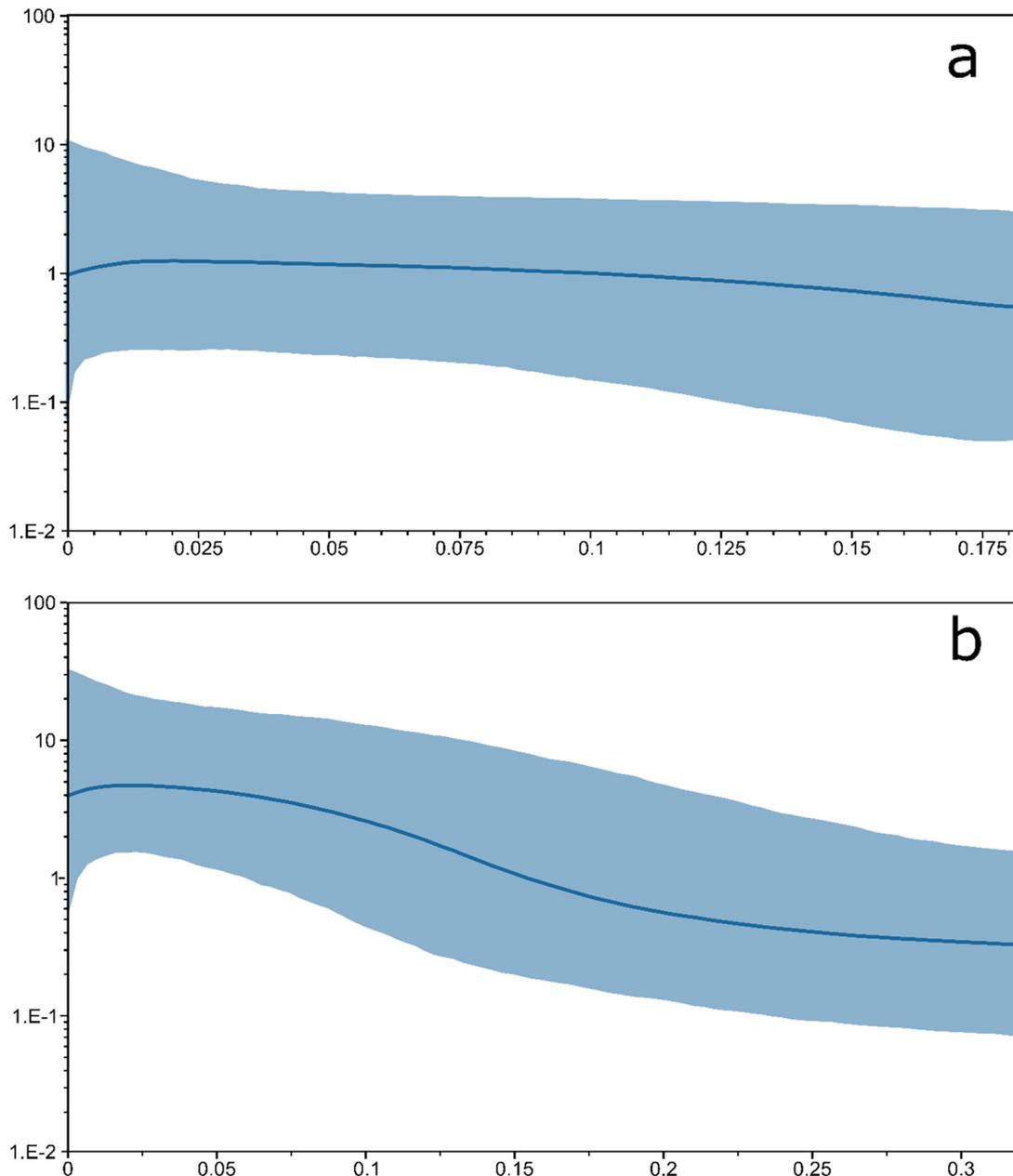


Fig. 2. Bayesian skyline plot representing historical demographic trends in: (a) *Pelasgus marathonicus*, and (b) *P. stymphalicus*. The course of the graph illustrates fluctuation of the population size from recent to the coalescence (left to the right). Time on the *x*-axis in million years ago. *Y*-axis shows effective population size \times generation time; in logarithmic scale. Solid middle line shows the median estimate and the blue shaded area represents 95% highest probability density limits.

The proportion of private haplotypes was noteworthy (14 out 18, *i.e.* 77.8%, Fig. 3). The private haplotypes were present in populations from all drainages with the exception of Kifissos (Athens). Within the native range, only the populations from Kifissos (Athens) and Erasinos shared one haplotype. The supposedly introduced population from Pinios (Thessaly) shared one haplotype with Kifissos (Beotia) and two with Sperchios populations. The only populations with a single haplotype were those from Assopos (Beotia) and Kifissos (Athens); both these populations occur in a limited part of very small isolated basins. Despite the fact that the

haplotypes from each population were closely related and grouped together, the haplotype network reconstruction did not show any broader geographic structure. DAPC showed that apart from the Assopos (Beotia) population, all *P. marathonicus* populations clustered together (Fig. S2). Genetic distances between populations ranged from 0.15% to 0.6% (Tab. 3). The highest genetic distance values were revealed between Assopos (Beotia) and Erasinos populations, Assopos (Beotia) and Kifissos (Athens) populations, despite the geographic proximity of these drainages, and Erasinos and Kifissos (Beotia). Genetic distances within groups were low

and ranged from zero (Kifissos-Athens and Assopos-Beotia) to 0.18% (Pinios-Thessaly) (Tab. 3). The mean genetic distance for the whole *P. marathonicus* dataset was 0.3%, while the maximum distance was 0.62%.

AMOVA showed that the genetic variance is evenly distributed among (52%) and between populations (48%) of *P. marathonicus* (Tab. 4), with a high and significant F_{ST} value

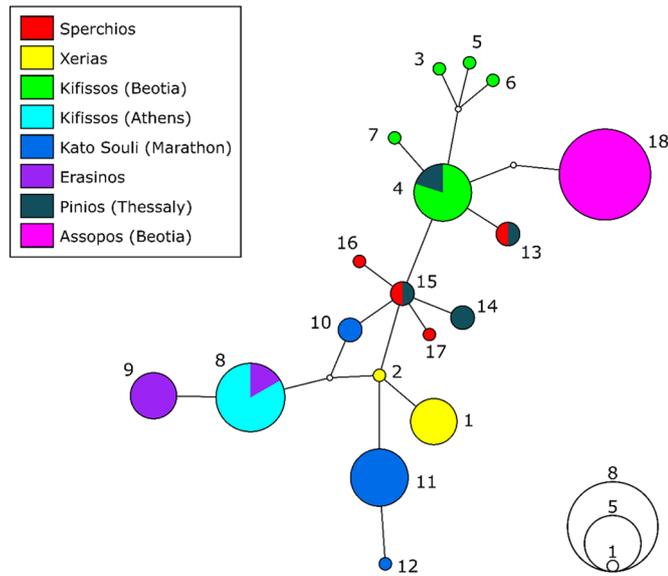


Fig. 3. Haplotype network reconstruction of *cytochrome b* sequences of *Pelasgus marathonicus* (statistical parsimony method, 95% connection limit). Numbers designate individual haplotypes. The maximum number of steps connecting parsimoniously two haplotypes is indicated (one line between two haplotypes being one mutational step). Missing intermediate haplotypes are shown as small white circles.

($F_{ST} = 0.52$), indicating a pronounced genetic differentiation of the populations. Pairwise F_{ST} s between populations also indicated a pronounced genetic differentiation (with the exception of the population pair Sperchios/Pinios-Thessaly), which was significant in half of the cases (Tab. 5).

3.2 *Pelasgus stymphalicus*

A total of 77 sequences of *cyt b* of 1140 bp length were obtained for *P. stymphalicus*. Our results revealed slightly higher nucleotide diversity within *P. stymphalicus* ($\pi = 0.0037$) than within *P. marathonicus*, and equal haplotype diversity ($Hd = 0.93$; Tab. 2). The number of obtained haplotypes was 31. All tests of neutrality were negative, of which Fu and Li's tests were significant (Fu and Li's $D = -2.39$, Fu and Li's $F = -2.47$). These results could indicate either population expansion or a selective sweep. Bayesian Skyline Plot reconstruction showed a gradual population growth over time, with a recent tendency to decrease, about 20 thousand years ago (Fig. 2b).

Haplotype network showed an interesting geographic pattern (Fig. 4). There were two haplogroups separated by a minimum of two mutations. The first haplogroup corresponded to the populations in the southern and eastern Peloponnese (Stymphalia Lake, Milos, Assopos-Peloponnese, Peristeras and Pamissos river basins), and the second corresponded to the populations on the mainland (Acheloos, Evinos and Mornos basins) and the adjacent north-western Peloponnese (Piros, Pinios and Kotichi basins). Interestingly, the haplotypes from Alfios were very divergent and were present in both haplogroups. The proportion of private haplotypes was very high (25 out of 31, *i.e.* 80.6%) and the private haplotypes were present in the populations from all drainages, with the exception of Pinios-Peloponnese population (a single haplotype shared with Kotichi population) and Pamissos (one

Table 3. Mean genetic distances for populations of *Pelasgus marathonicus* based on *cytochrome b* sequences (uncorrected p-distances), in %. Within group distances are on diagonal in bold.

	Xer	KifB	Era	Kat	KifA	Spe	Pin	Ass
Xerias	0.035							
Kifissos (Beotia)	0.32	0.135						
Erasinos	0.32	0.5	0.035					
Kato Souli (Marathon)	0.19	0.33	0.32	0.135				
Kifissos (Athens)	0.25	0.43	0.07	0.25	0			
Sperchios	0.25	0.21	0.42	0.25	0.35	0.175		
Pinios (Thessaly)	0.25	0.18	0.42	0.25	0.35	0.15	0.140	
Assopos (Beotia)	0.42	0.25	0.6	0.43	0.53	0.31	0.28	0

Table 4. Partitioning of genetic variance, in %, and fixation indices (F_{ST}) for *Pelasgus marathonicus* and *P. stymphalicus*, based on *cytochrome b* sequences.

		<i>P. marathonicus</i>	<i>P. stymphalicus</i>
Partitioning of genetic variance	Among populations	52.11	58.34
	Within populations	47.89	41.66
Genetic differentiation	F_{ST}	0.521	0.583

Table 5. Pairwise F_{ST} between populations of *Pelagus marathonicus* based on *cytochrome b* sequences. Significant values (at $\alpha = 0.05$ /number of pairs) indicated in bold.

	Xer	KifB	Era	Kat	KifA	Pin	Spe
Xerias							
Kifissos (Beotia)	0.703						
Erasinos	0.889	0.806					
Kato Souli (Marathon)	0.508	0.591	0.706				
Kifissos (Athens)	0.929	0.805	0.750	0.677			
Pinios (Thessaly)	0.643	0.246	0.792	0.457	0.800		
Sperchios	0.603	0.276	0.77	0.405	0.781	0	
Assopos (Beotia)	0.969	0.733	0.979	0.843	1.000	0.809	0.815

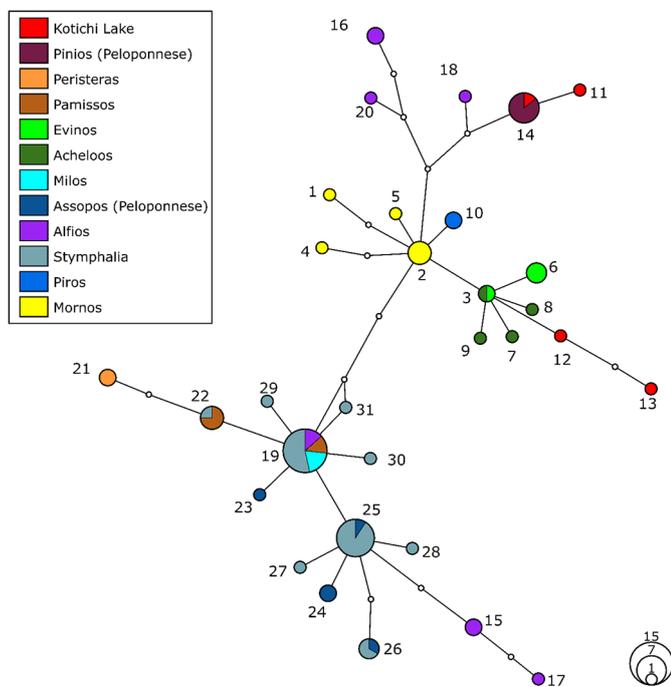


Fig. 4. Haplotype network reconstruction of *cytochrome b* sequences of *Pelagus stymphalicus* (statistical parsimony method, 95% connection limit). Numbers designate individual haplotypes. The maximum number of steps connecting parsimoniously two haplotypes is indicated (one line between two haplotypes being one mutational step). Missing intermediate haplotypes are shown as small white circles.

haplotype shared with Stymphalia and the other with Stymphalia, Alfios and Milos populations). In the north-western haplogroup two haplotypes were shared between two different populations: Evinos/Acheloos and abovementioned Pinios-Peloponnese/Kotichi; in the south-eastern haplogroup, four haplotypes were shared (see Fig. 4). There was a particular geographical pattern within the north-western haplogroup, where the haplotypes of the pairs of populations grouped together: Piros (Peloponnese) and Mornos (mainland), Evinos and Acheloos (both mainland) and haplotypes 11 and 14 from Kotichi and Pinios (both Peloponnese). The haplotypes from

Kotichi population were diversified: two other haplotypes (12 and 13) were most similar to the Evinos/Acheloos group (mainland). This could indicate a secondary contact of the populations from the mainland and north-western Peloponnese during the Pleistocene. Within the south-eastern haplogroup there was no geographic pattern.

DAPC divided *P. stymphalicus* populations into four clusters: (1) Kotichi and Pinios-Peloponnese, (2) Evinos and Acheloos, (3) Mornos and Piros populations and (4) the remaining populations (Fig. S3). These results line up with those of the haplotype network reconstruction (Fig. 4). The mean genetic distances (p -distances) between populations ranged from 0.1% to 0.8%, being the highest between Peristeras and Kotichi and Pinios-Peloponnese populations and the lowest for the pairs of populations Acheloos/Evinos, Milos/Parnissos, Stymphalia/Parnissos, Stymphalia/Milos (Tab. 6). In general, the high p -distances were detected between populations from one haplogroup *versus* populations from the other haplogroup. Genetic distances within populations ranged from 0 (Piros, Pinios-Peloponnese, Milos) to 0.5% (Alfios basin), see Table 6. A high within population distance was also observed for Kotichi (0.42%). The mean genetic distance for the whole *P. stymphalicus* dataset was 0.4% and the maximum difference was 0.89%.

AMOVA showed that genetic variance in *P. stymphalicus* was higher among (58%) than between populations (42%) and the F_{ST} of 0.58 indicated a pronounced genetic differentiation of the populations (Tab. 4). Also pairwise F_{STs} indicated a pronounced genetic differentiation between the populations (with the exception of Assopos-Peloponnese *vs.* Parnissos and Milos), but were, however, significant only for most comparisons of Assopos-Peloponnese and two of Mornos populations (Tab. 7).

4 Discussion

Both investigated *Pelagus* species have particular distribution patterns (Fig. 1), probably given not only by the geological history of the area but also by their ability to inhabit very small streams, isolated spring-fed lentic waters and wetlands, unlike most other native fish species in the area. *Pelagus* generally inhabits slow-flowing and nearly stagnant waters with abundant vegetation from coastal wetlands to upland plateaus. It is able to survive periods of draught in river-

Table 6. Mean genetic distances between populations of *Pelagus stymphalicus* based on *cytochrome b* sequences (uncorrected p-distances), in %. Within group distances are on diagonal in bold.

	Mor	Evi	Ach	Pir	Kot	Pin	Alf	Per	Pam	Mil	Ass	Sty
Mornos	0.125											
Evinos	0.22	0.044										
Acheloos	0.22	0.13	0.132									
Piros	0.15	0.24	0.24	0								
Kotichi	0.35	0.35	0.35	0.37	0.424							
Pinios (Peloponnese)	0.33	0.42	0.42	0.35	0.29	0						
Alfios	0.43	0.53	0.52	0.46	0.6	0.52	0.497					
Peristeras	0.59	0.68	0.68	0.61	0.81	0.79	0.62	0				
Pamissos	0.38	0.47	0.47	0.40	0.60	0.58	0.41	0.21	0.053			
Milos	0.33	0.42	0.42	0.35	0.55	0.53	0.36	0.26	0.05	0		
Assopos (Peloponnese)	0.48	0.58	0.58	0.51	0.71	0.68	0.47	0.42	0.21	0.16	0.193	
Stymphalia	0.40	0.49	0.49	0.43	0.63	0.60	0.41	0.34	0.13	0.08	0.16	0.112

Table 7. Pairwise F_{ST} between populations of *Pelagus stymphalicus* based on *cytochrome b* sequences. Significant values (at $\alpha = 0.05/\text{number of pairs}$) indicated in bold.

	Mor	Evi	Ach	Pir	Kot	Pin	Alf	Per	Pam	Mil	Ass
Mornos											
Evinos	0.367										
Acheloos	0.564	0.866									
Piros	0.409	0.611	0.333								
Kotichi	0.793	1.000	0.960	0.879							
Pinios (Peloponnese)	0.288	0.229	0.333	0.208	0.369						
Alfios	0.823	1.000	0.952	0.857	1.000	0.626					
Peristeras	0.246	0.188	0.37	0.314	0.453	0.215	0.377				
Pamissos	0.727	1.000	0.937	0.814	1.000	0.553	1.000	0.114			
Milos	0.682	0.71	0.778	0.711	0.873	0.581	0.653	0.221	0.265		
Assopos (Peloponnese)	0.714	0.763	0.797	0.767	0.853	0.73	0.703	0.366	0.072	0.091	
Stymphalia	0.749	0.897	0.896	0.814	0.96	0.638	0.806	0.251	0.388	0.417	0.291

bed pools, wells and springs (Economou *et al.*, 1999; Economou, 2000) and although it prefers freshwater, it can potentially survive for short periods in brackish conditions with salinity up to 12–13‰ (Bianco and Nordlie, 2008). Their ecology and ability to stand adverse conditions most probably enabled *P. marathonicus* and *P. stymphalicus* to have a wide distribution in seasonally semi-arid areas.

The range of *P. stymphalicus* was suggested to include most of the Ionian and a few Aegean drainages on the Peloponnese, including endorheic Stymphalia Lake (type locality), and the rivers in the south-west of mainland, in the Ionian Sea slope, from the Mornos to two small drainages in the south-eastern part of the Amvrakikos Gulf, including Lefkada Island (Kottelat and Barbieri, 2004; Economou *et al.*, 2007; Koutsikos *et al.*, 2012; Barbieri *et al.*, 2015). However, many of these populations were ascribed only tentatively to *P. stymphalicus*, and the verification of their conspecific status has been awaiting further studies (Kottelat and Freyhof, 2007). We confirmed the presence of *P. stymphalicus* in the majority of these drainages. Moreover, we have found it for the first time in the Assopos (north-eastern Peloponnese), where earlier surveys failed to find it (Economou *et al.*, 1999, 2007). To summarize, based on our results, *P. stymphalicus* occurs in the

Ionian region, from the Acheloos basin on the mainland to the Pamissos river in the south-western Peloponnese, and in the eastern Peloponnese (Aegean Sea slope), in the lowland Argolid’s spring-fed wetlands (in our samples represented by Milos population). However, the status of a small number of populations along the Akarnanian coast was not possible to examine (Astakos, Lefkada, small drainages in the south-eastern part of the Amvrakikos Gulf), so the exact distribution limit at the north-western edge of the distribution range is still undefined. Moreover, the identity of *Pelagus* from the endorheic Kandila springs in the north-east of the Peloponnese (adjacent to the Stymphalia drainage), where another species has been assumed to be present (Kottelat and Barbieri, 2004), needs further investigation.

Furthermore, we were not able to confirm the occurrence of *Pelagus* in several small drainages in the Peloponnese which lie within the range of the species, despite our sampling effort (*e.g.* Neda, Vouraikos, Selinous, Taka Lake). Importantly, in several small basins in the Peloponnese (Peristeras, Assopos, Piros) this species was found in very low numbers, which may indicate that it is under threat.

A particular situation in the Alfios basin on the Peloponnese, revealed by Kottelat and Barbieri (2004), was

confirmed: the populations from the lower part of the basin belong to *P. stymphalicus*, whereas the populations from the headwater part belong to *P. laconicus* (genetic divergence between these species 10.2%, own data). This is probably a result of past tectonic or orogenetic events, which led to river captures in headwaters of the Alfios and Evrotas river basins, which, at the same time separated the population of *P. laconicus* currently occurring in the upper Alfios from its remaining populations in the Evrotas river. Furthermore, the diversity and divergence of haplotypes within the Alfios population of *P. stymphalicus* may indicate past river capture events which enabled a connection of the contemporary tributaries of Alfios, which had previously belonged to different drainages and hosted genetically distant populations. The affinity of the populations of *P. stymphalicus* from Evinos, Acheloos and Kotichi Lake, and the populations from Mornos and Piros drainages (Fig. 4) can be explained by the Pleistocene sea level regressions (Perissoratis and Conispoliatis, 2003). Such regressions allowed the connection of the rivers on the mainland and on the Peloponnese, a phenomenon that enabled contact between the populations. This is in accordance with a very similar pattern that has been observed in the distribution of other freshwater fish species inhabiting this region (e.g. *Luciobarbus albanicus* (Steindachner 1870) or *Tropidophoxinellus hellenicus* (Stephanidis 1971), Economou *et al.*, 2007). Acheloos and Evinos populations of *Pelaspus* appear to be closely related, and even share one haplotype (Fig. 4). Durand *et al.* (1999) suggested a connection of both drainages during the Late Pleistocene, and our results support that hypothesis. The divergence of the haplotypes from Kotichi can indicate a secondary contact between the population from the Peloponnese (haplotypes 14 and 11 grouped with the haplotype from the Pinios population) and the mainland (haplotypes 12 and 13 were similar to the Evinos/Acheloos group, see Fig. 4).

Analysed samples of *Pelaspus marathonicus* cover almost all known populations of the species, and include specimens from the type locality, Kato Souli in Schinias-Marathon wetland in Attiki. We confirmed that all investigated populations, which were previously supposed to belong to this species, although some of them only tentatively (Kottelat and Barbieri, 2004; Economou *et al.*, 2007; Koutsikos *et al.*, 2012; Barbieri *et al.*, 2015), are conspecific. The species range spans from the Erasinios drainage, just south of Athens, up to the Xerias basin in the Pagasitikos Gulf. *Pelaspus marathonicus* was found even in the Beotian Assopos drainage, where it had been considered possibly extirpated (Economou *et al.*, 2007). Furthermore, a population of *Pelaspus* from the Thessalian Pinios river drainage, considered to be recently introduced (Koutsikos *et al.*, 2012), was shown to be *P. marathonicus*. We were not able to trace the exact origin of the introduced population, because haplotypes present in the Pinios population are shared with both Sperchios and Kifissos (Beotia) populations. More samples from all basins should be tested to reveal the path of introduction.

The only known population of *P. marathonicus* which we were not able to analyse is the one from central-western Euboea Island (reported by Barbieri *et al.*, 2015). Although this population is virtually within the range of the species, its identity needs further research, as other freshwater fishes from Euboea in fact show affinity not to the nearby occurring

species, but to populations from geographically distant areas (*Barbus* Daudin 1805 to Ionian populations: Tsigenopoulos and Berrebi, 2000, see further discussion; *Squalius* to eastern Aegean populations: Zardoya and Doadrio, 1999). Further, *Pelaspus* may occur in some very small unsampled basins.

4.1 Biogeographical insights

Several biogeographic regions were recognised in the southern Balkans based on the distribution of fish species (Zogaris *et al.*, 2009; Oikonomou *et al.*, 2014). Southern Greece comprises two such regions, Ionian and Western Aegean. The extent of both regions slightly varies according to different authors (Zogaris *et al.*, 2009; Oikonomou *et al.*, 2014; Zogaris and Economou, 2017; Vavalidis *et al.*, 2019). Recent molecular genetic studies conducted on several fish species suggest that further subdivision of the Ionian region is possible, dividing it to southern and northern subregion, with a boundary immediately south of the Amvrakikos Gulf (see overview in Buj *et al.*, 2019). A part of the ichthyofauna of the southern Ionian subregion actually shows affinities to that of the Western Aegean region. *Pelaspus stymphalicus* (southern Ionian region) and *P. marathonicus* (Aegean region) are a pair of sister species (Schönhuth *et al.*, 2018). The same relationship was found in other primary freshwater fishes from these regions, i.e. in the genus *Rutilus* Rafinesque 1820 (Ketmaier *et al.*, 2008; Perea *et al.*, 2010; Geiger *et al.*, 2014; Schönhuth *et al.*, 2018), *Telestes* Bonaparte 1840 (Perea *et al.*, 2010; Geiger *et al.*, 2014; Buj *et al.*, 2017; Schönhuth *et al.*, 2018) and for some species of *Barbus* (*Barbus peloponnesius* Valenciennes 1842 and *Barbus euboicus* Stephanidis 1950, see Tsigenopoulos and Berrebi, 2000). On the other hand, species of other genera present in both areas are not closely related. This applies to *Luciobarbus* Heckel 1843 (Yang *et al.*, 2015), *Squalius* (Perea *et al.*, 2010) and *Scardinius* Bonaparte 1837 (Ketmaier *et al.*, 2004; Perea *et al.*, 2010). These differences in the evolutionary relationships within different genera are most probably a consequence of a very complex geological history of the area, in a combination with the different life history traits of the species/genera and repeated episodes of colonization of the area.

The genetic diversity of the endemic fishes from this area was studied only in a few species and showed different patterns than the ones revealed here for *Pelaspus*. For the Western Aegean region, no comparative study of such extent is available. *Pelaspus marathonicus* is the most widespread fish species in this region, having a much broader distribution than any other species occurring here. Moreover, although most of its populations have private haplotypes, these are closely related. On the contrary, other endemic species from the Western Aegean region have much smaller distribution ranges, they occur in a single (i.e. *Alburnoides economou* Barbieri, Vukić, Šanda & Zogaris 2017, *Rutilus ylikiensis* Economidis 1991, *Scardinius graecus* Stephanidis, 1937) or a few drainages (i.e. *Luciobarbus graecus* (Steindachner 1895), *Telestes beoticus* (Stephanidis 1939)) (Economou *et al.*, 2007; Barbieri *et al.*, 2015, 2017), and their genetic diversity has not been studied yet.

Pelaspus stymphalicus is also a widespread species. None of the other species from the southern part of the Ionian region has such a large range (Economou *et al.*, 2007). A recent study

has revealed a high genetic diversity in *T. pleurobipunctatus*, a species that has been previously considered a single species widely distributed in the whole Ionian region. In fact, *T. pleurobipunctatus* comprises several distinct evolutionary units which probably represent distinct species (Buj *et al.*, 2019). Within its overlapping range with *P. stymphalicus*, three evolutionary units of *T. pleurobipunctatus* occur. Interestingly, there is a similar pattern in the close relationships between *Telestes* populations from the mainland (Evinos river) and the north-western Peloponnese (Kotichi and Pinios basins) to the one observed in *Pelasgus*, suggesting a connection of these areas during the Pleistocene glacial periods (Buj *et al.*, 2019). On the other hand, while the populations of *Pelasgus* from Acheloos and Evinos are genetically related and even share one haplotype, the populations of *Telestes* from these drainages form two distinct clades, differing on *cyt b* by 3.5% (Buj *et al.*, 2019). Also *Squalius* displays higher diversity in the overlapping range with *P. stymphalicus*. Durand *et al.* (1999) observed differentiation between *Squalius* from Alfios and a clade which includes *Squalius* from Peloponnesian Pinios and mainland Evinos and Acheloos. Further, Perea *et al.* (2016) found divergence between *Squalius* from Alfios and Stymphalia Lake. Thus *P. stymphalicus* seems to be unique in its genetic homogeneity over a large area, compared with other widespread fish species.

4.2 Threats and conservation

The IUCN status of *P. marathonicus* is near threatened (NT), while that of *P. stymphalicus* is least concern (LC). However, our results suggest a recent decreasing tendency of the effective population sizes in both studied species. This suggests that populations of both species could be sensitive to anthropogenic pressure. Main threats to freshwater ecosystems are habitat loss, introduction of alien species and pollution (Collen *et al.*, 2014). The human-induced hydrological degradation, construction of dams and other dispersal barriers and canalization of water bodies entail the destruction of the habitat of *Pelasgus* species. A spread of alien species, especially *Gambusia holbrooki* Girard 1859, has long been shown to be a serious problem for this and other small-sized endemic fishes in Greece (Economidis 2009; Kalogianni *et al.*, 2019). Furthermore, since fish translocations are on the rise (Koutsikos *et al.*, 2019), an insidious threat is hybridization with other *Pelasgus* species, which can be traced by the use of appropriate nuclear markers. Cases of hybridization due to translocation were already observed in some *Pelasgus* populations (own unpublished data). *Pelasgus* is not a species of interest for recreational fishing, neither is it used as live bait. Its translocation is unintentional, and happens by translocation of species of commercial or recreational interest, with which individuals of *Pelasgus* are mixed (Kottelat and Barbieri, 2004).

Although distribution ranges of the studied species seem to be relatively large, *Pelasgus* are small-sized fishes with limited dispersal abilities, and given the type of habitat they inhabit (usually small lowland streams, with slow-flowing water), they are susceptible also to droughts caused by regional climate variation and global climate change. Climate change predictions forecast increased desiccation of perennial flowing

water bodies and wetlands in the region (Kokkoris *et al.*, 2019).

It is alarming that *Pelasgus* populations were not found in our and other recent surveys in several river basins that lie within their distribution ranges (e.g. Economou *et al.*, 2016; Zogaris *et al.*, 2018). The presence of private *cyt b* haplotypes in populations of both species points to the uniqueness and vulnerability of each population. Populations with lower genetic diversity could have less chances of survival due to a small effective population size, and due to a smaller capacity to adapt to changing environmental conditions (Buj *et al.*, 2015). It is of great importance to protect the species as a whole, but it should not be overlooked that units below species level may be of exceptional relevance for the future survival of the species. As the genetic structure of all populations from each river basin is distinct, populations should be considered individually. A strict protection of the rivers, lowland springs and wetlands where *P. stymphalicus* and *P. marathonicus* occur is of a great importance. The conservation status of both species should be re-evaluated. Of special conservation relevance is the Western Aegean ecoregion, being already considered a species-poor and with many perennial river reaches and wetlands under threat.

Supplementary Material

Supplementary Tables S1 to S3.

The Supplementary Material is available at <https://www.kmae-journal.org/10.1051/kmae/2020019/olm>.

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